

Climate and human-driven variability of summer hypoxia on a large river-dominated...

[Health & Medicine](#)



Introduction

Coastal hypoxia has become common especially in large river dominated coastal ecosystems, where a substantial amount of oxygen is consumed in subsurface waters via oxidation of organic matter coming from either external sources or internal primary production ([Breitburg et al., 2018](#)). Recent studies show hypoxia has become more wide-spread or severe in the last two decades; the number of reported global “ dead zones” is increasing ([Breitburg et al., 2018](#)), the area of hypoxia is expanding, and the minimum values of dissolved oxygen (DO) concentration are decreasing, even approaching anoxia. However, using a single property, such as hypoxic area, may be biased and insufficient to elucidate the severity of hypoxia. Other properties, such as thickness and minimum and average oxygen concentrations, are also crucial for exploring the severity of the hypoxia problem ([Matli et al., 2018](#) ; [Scavia et al., 2018](#)).

Climate change and human activities influence coastal hypoxia via immediate or direct drivers and peripheral or indirect drivers. By immediate or direct drivers, we mean those factors that consume O₂ or limit the supply of O₂ to the bottom water ([Fennel and Testa, 2018](#)). The former includes water column and benthic organic carbon respiration ([Resplandy, 2018](#)), while the latter includes vertical advection and diffusion ([Zhu et al., 2016](#) ; [Markus Meier et al., 2018](#)) (as characterized by water column stratification), lateral advection ([Fennel and Testa, 2018](#)) [as characterized by subsurface water residence time (τ)], and oxygen solubility ([Lui et al., 2014](#)). By peripheral or indirect factors, we mean those that lead to the direct factors.

For example, eutrophication or high primary production in surface waters can lead to high respiration and thus hypoxia in bottom waters, and temperature and/or salinity gradients can lead to water column stratification. Most of the research on the coastal hypoxia mechanisms has focused on indirect factors, such as eutrophication driven by human activity ([Rabalais et al., 2014](#) ; [Wang et al., 2016](#)) and ensuing enhancement of primary production and organic matter input ([Chen et al., 2007](#) ; [Li et al., 2018](#)), or temperature and non-temperature effects driven by climate change ([Altieri and Gedan, 2015](#) ; [Li et al., 2016a](#)). We argue that the relative contribution of each of these factors is still not clear. Because these indirect factors influence hypoxia via direct pathways, we must disentangle the contributions of the direct drivers in order to quantify the contributions of climate change and human activities and to delineate the pathways impacting the long-term variation of coastal hypoxia.

The relative contributions of direct drivers controlling long-term hypoxia also vary due to location and climatological regimes of local marine ecosystems. In this study, we have developed a hypoxia index (HI) as defined by the first component of principal component analysis (PCA), using data of observable properties from eight cruises in the Changjiang plume-impacted shelf area in the East China Sea (ECS). We then use a regression analysis of the derived HI and three direct hypoxia drivers (stratification, τ , and biological respiration) to quantify their relative contributions to hypoxia severity. As these direct factors were not available for long-term trend analysis, we used long-term observed data and model results, to reconstruct the HI and to explore the contribution of climate change (associated with stratification, τ , <https://assignbuster.com/climate-and-human-driven-variability-of-summer-hypoxia-on-a-large-river-dominated-shelf-as-revealed-by-a-hypoxia-index/>

and oxygen solubility) and human activities (associated with biological respiration) to variation in hypoxia over the past 60 years. This study also has implications for understanding hypoxia in estuaries and bays affected by substantial nutrient runoff and climate change.

Material, Model, and Methods

Study Area and Sampling Collection

Changjiang River (also Yangtze River) is the largest river in the Euro-Asian continent, freshening the area continuously ([Park et al., 2011](#)) with its annual discharge of $924.8 \times 10^9 \text{ m}^3 \text{ yr}^{-1}$ ([Liu et al., 2016](#)). Changjiang plume-impacted shelf area is one of the most productive continental shelf sea in the world due to the mixing of nutrient-laden Changjiang River plume and Kuroshio Current. Summer hypoxia occurs with strong stratification and prolonged residence time for subsurface water ([Zhang et al., 2019](#)), and also fueled by marine organic matter exporting from the surface ([Wang et al., 2016](#)). Hypoxia has been recorded since 1950s, and has become more severe recently with DO minimum decreasing and hypoxic area expanding ([Zhu et al., 2011](#) ; [Wei et al., 2017](#)).

We visited the 31°N transect in the ECS eight times in August 2006, August 2009, September 2010, August 2011, August 2013, August 2014, August 2016, and August 2017 ([Supplementary Figure 1](#)). For this study, we focused on an area between 30.5°N and 32°N, 122.5°E and 125°E, and assembled oceanographic data over the past 60 years. This is the only transect in the region where long-term data are available.

At each station, temperature and salinity were measured with a Sea Bird model 911 conductivity-temperature depth (CTD) recorder. Water samples for measuring concentrations of nutrients, chlorophyll *a* (Chl *a*), and DO were collected with Niskin bottles attached to the CTD rosette.

Chemical Analysis

Concentrations of nutrients (NO_3^- , NO_2^- , NH_4^+ , PO_4^{3-} , and SiO_3^{2-}) were determined with colorimetric methods, and DO was measured by Winkler titration method ([Grasshoff et al., 1999](#)). Chl *a* was measured using a Turner Designs 10-AU fluorometer ([Holm-Hansen et al., 1965](#)).

Conceptual Model

Observable Properties for Extended Definition of Hypoxia Severity

Since the status of hypoxia is not merely reflected by the area or oxygen concentration, it is necessary to find an index to express the hypoxia severity more comprehensively. In this study, we derive the index based on a PCA of six observable properties ([Supplementary Table 1](#)): sectional area, mean thickness, the longitude span of the hypoxia area ([Figure 1](#)), DO minimum, mean bottom DO, and mean DO concentration when saturation (DO%) was below 100%. Since the conventional hypoxia threshold of 2 mg L^{-1} is below the empirical sublethal and lethal DO thresholds for many organisms ([Vaquer-Sunyer and Duarte, 2008](#)), we defined hypoxia area as the vertical sectional area surrounded by the $94 \text{ } \mu\text{mol L}^{-1}$ (3 mg L^{-1}) contour along the 31° N transect, the same threshold for hypoxia as previous reports defined ([Chen et al., 2007](#); [Levin et al., 2009](#); [Zhou et al., 2017](#)).

The mean thickness of the hypoxia area (D_{hy} , m) is the averaged vertical

<https://assignbuster.com/climate-and-human-driven-variability-of-summer-hypoxia-on-a-large-river-dominated-shelf-as-revealed-by-a-hypoxia-index/>

distance between the $94 \mu\text{mol L}^{-1}$ contour and the bottom of the stations. The longitude span of the hypoxia area is the maximum distance of the $94 \mu\text{mol L}^{-1}$ contour along 31°N transect.

FIGURE 1

A conceptual model for the dimension of sectional hypoxia area and water and salt budget. V_Q , volume of Changjiang River runoff; V_g , volume of groundwater runoff; V_z , vertical exchange volume between layer 1 and 2; V_{ent} , volume of advection-induced entrainment; V_s , surface advective outflow volume from layer 1; V_d , deep advective inflow volume to the layer 2; S_1 , salinity of layer 1; S_2 , salinity of layer 2; S_{1oc} and S_{2oc} , are oceanic endmember salinities of layer 1 and 2, respectively.

Calculating the Direct Drivers Using the LOICZ Model

In order to quantify the contributions of the direct drivers in regulating hypoxia, we calculated τ and respiration rates during each cruise, using a two-layer river dominated LOICZ (Land-Ocean Interactions in the Coastal Zone) model ([Figure 1](#)), which was based on water and salt balances within each layer at steady state ([Gordon et al., 1996](#)). Layers 1 and 2 were divided by the mixed layer depth (MLD). We assumed the Changjiang River and the Kuroshio Current are the two major water masses mixing in the studied system. As groundwater data is very limited and the distribution is patchy, we assumed fresh groundwater only contribute to the layer 1 and a constant flux (V_g) of $0.02 \text{ m}^3 \text{ m}^{-2} \text{ d}^{-1}$ ([Guo et al., 2020](#)) with average

salinity (S_g) of 5 and phosphate of $3.3 \mu\text{mol L}^{-1}$ ([Wang et al., 2018](#)). Thus, water and salt balances in the entire system are:

$$V_s (\text{output}) = V_Q + V_g + V_d (\text{inputs}) \quad (1)$$

$$V_s \times S_1 (\text{output}) = V_d \times S_{2oc} + V_Q \times S_Q + V_g \times S_g (\text{inputs}) \quad (2)$$

The salt contributed by Changjiang River water and groundwater were far less than deep advective inflow (V_d) [salinity of Changjiang River water (S_Q) at the Xuliujing station ([Supplementary Figure 1](#)) was usually <0.15 (unpublished data)] and the total discharge of Changjiang River water and groundwater was far less than V_d , then the last two items ($V_Q \times S_{2Q} + V_g \times S_{2g}$) of Eq. 2 could be ignored.

A combination of Eqs1 and 2 gives:

$$V_d = (V_Q + V_g) \times S_1 / (S_{2oc} - S_1) \quad (3)$$

For the bottom layer, the salt balance is:

$$V_d \times S_{2oc} = V_{ent} \times S_2 \quad (4)$$

The variables are defined in [Figure 1](#).

Stratification ($\Delta\sigma$) was defined as the average difference in the density anomaly between bottom and surface (0–2 m depth) at stations along the transect. We used the simple linear equation of seawater state ([Intergovernmental Oceanographic Commission \[IOOC\], 2010](#)) to estimate the density difference between bottom and surface seawater,

$$\Delta\sigma = \rho_b (S_b, T_b, p_b) - \rho_s (S_s, T_s, p_s) \quad (5)$$

<https://assignbuster.com/climate-and-human-driven-variability-of-summer-hypoxia-on-a-large-river-dominated-shelf-as-revealed-by-a-hypoxia-index/>

where T_s and T_b are surface and bottom temperature ($^{\circ}\text{C}$), S_s and S_b are surface and bottom absolute salinity, and p_s and p_b are surface and bottom pressure (dbar), respectively. For reducing the observational errors of S_s and S_b , we used historical data from the Japan Meteorological Agency, Oceanographic Data Center (JODC), the Korea Oceanographic Data Center (KODC), and the World Ocean Atlas (WOA), which were averaged in the specified box (122.5°E - 125°E , 30.5°N - 32°N , [Supplementary Figure 1](#)), besides of our own data. MLD is defined as the depth at which the density differs by 0.125 from the surface value ([Huang and Russell, 1994](#)).

Average water residence time is typically expressed as the ratio of system volume to the average rate of outflow through the system ([Gordon et al., 1996](#)). For layer 2, there are two types of water mixing; one is the vertical exchange with layer 1 including diffusion and advection (including entrainment) (V_z), and the other is layer 2 water exchange with the offshore subsurface water (V_d). The latter often accounts for most of the water mixing. If so, then τ is the total volume of subsurface water divided by the water inputs (outputs) or V_2/V_d , can be simply presented by combining Eq. (3) for canceling out V_d :

$$\tau = V_2 / V_d = V_2 \times (S_2 - S_1) / (V_Q + V_g) \quad (6)$$

The long-term τ can be calculated based on Eq. 6, V_2 was obtained by multiplying the depth below MLD and assuming a fixed study area of $1.8 \times 10^{10} \text{ m}^3$ along the transect (122°E - 124°E , 30.5°N - 31.5°N).

The last direct driver is the respiration rates in the water (R_h , $\text{mmol C m}^{-2} \text{d}^{-1}$), which we derive from the difference between NPP (net primary production) and NEP (net ecosystem production), as $\text{NPP} = \text{GPP} - R_a$ and $\text{NEP} = \text{GPP} - (R_a + R_h)$. GPP is gross primary production, and R_a is autotrophic respiration rates. NPP data were acquired from the Marine Satellite Data Online Analysis Platform (SatCO₂) [1](#), including monthly SeaWiFS VGPM NPP from 1997 to 2007 and MODIS Aqua VGPM NPP from 2002 to 2016. These NPP products were corrected by euphotic layer depth including the effect of colored dissolved organic matter absorption ([He et al., 2017](#)). NEP was calculated using the LOICZ model, as the net biological uptake or release of carbon calculated from phosphate stoichiometrically ($\text{C/P} = 106$) ([Gordon et al., 1996](#)) within each layer.

Thus, we have,

$$R_h = \text{NPP} - \text{NEP} \quad (7)$$

R_h of $37.0 \text{ mmol C m}^{-2} \text{d}^{-1}$ from *in situ* incubation was applied for 2014 due to too much cloud for deriving satellite NPP in the study area. The mean values and uncertainties (standard deviation) of satellite NPP and LOICZ model derived NEP are 68.8 ± 7.9 and $28.2 \pm 68.9 \text{ mmol C m}^{-2} \text{d}^{-1}$, respectively, based on the eight cruises with an error propagation method ([Taylor, 1997](#)). The mean value and uncertainty of R_h is estimated to be $44.7 \pm 70.2 \text{ mmol C m}^{-2} \text{d}^{-1}$. The volumetric respiration rates ($\text{mmol C m}^{-3} \text{d}^{-1}$) was assumed constant over the whole water column in the hypoxia area. This approach is verified and deemed acceptable based on our results during

2011 cruise at five stations of 24 h-incubation for respiration along the transect in Changjiang plume-impacted shelf area (averaged 30.1 ± 24.8 , 35.5 ± 19.2 , and 26.4 ± 21.5 for 2 m, 10 m, and bottom waters, respectively). The integrated respiration rates in the hypoxia area (R_{hy} , $\text{mmol C m}^{-2} \text{d}^{-1}$) depends on the thickness of the hypoxia area and the water depth (D , m) over longitudes (L) spanning the sampling stations from west (L_w) to east (L_e):

$$R_{hy} = R_h \int_{L_w}^{L_e} (D_{hy} / D) dL \quad (8)$$

The contribution of sediment to R_{hy} was assumed to be 50% ([Zhou et al., 2017](#)) and an average $8.43 \text{ mmol O}_2 \text{ m}^{-2} \text{d}^{-1}$ of sediment oxygen consumption ([Song et al., 2016](#)) compare to our $16.26 \text{ mmol O}_2 \text{ m}^{-2} \text{d}^{-1}$ of R_{hy} in 2011. Redox-mediated phosphorus desorption from inorganic particles were assumed to be very limited due to adsorption-desorption equilibrium at mineral surfaces ([Meng and Yao, 2014](#)) and very high phosphorus burial efficiency ([Liu et al., 2016](#)) in the study area.

Long-Term Data

Changjiang River discharge (Q) data were downloaded from the website of the Ministry of Water Resources of the People's Republic of China. [2](#) The discharges were measured at the hydrological gauge station, Datong, which has no tidal influence and is 624 km upstream from the river mouth.

Dissolved inorganic nitrogen (DIN) and dissolved inorganic phosphate (DIP) fluxes in August were calculated as the product of Changjiang River discharge in August and nutrient concentrations at Datong gage station or

<https://assignbuster.com/climate-and-human-driven-variability-of-summer-hypoxia-on-a-large-river-dominated-shelf-as-revealed-by-a-hypoxia-index/>

Xuliujing station ([Supplementary Figure 1](#)) in historical records ([Gu, 1980](#) ; [Quan et al., 2005](#) ; [Wang, 2006](#) ; [Li et al., 2007](#) , [2016b](#) ; [Dai et al., 2011](#) ; [Gao et al., 2012](#)) and *in situ* observations during summers of 2009, 2010, 2011, 2013, 2014, and 2016 at Xuliujing station.

Multivariate ENSO index (MEI) was obtained from <https://www.esrl.noaa.gov/psd/enso/mei/> ([Wolter and Timlin, 1998](#)).

Sea surface temperature (SST) in August along the 31°N transect was provided by the NOAA-ESRL Physical Sciences Division, Boulder Colorado from their website at <https://www.esrl.noaa.gov/psd/> , named as COBE SST2.

In situ observations of long-term temperature and salinity in August within the study box were from JODC, [3](#) KODC, [4](#) and WOA version 2013. [5](#)

Wind speed data were acquired from the International Comprehensive Ocean-Atmosphere Data Set (ICOADS) with 2-degree spatial resolution at the website of the Royal Netherlands Meteorological Institute (KNMI [6](#)).

Data Analysis

To quantify the hypoxia severity, we applied PCA to the six observable properties. PCA is a useful technique for reducing the dimensionality of large datasets, increasing interpretability but at the same time minimizing information loss. Numerous variables including those correlated with each other can be synthesized to create new uncorrelated variables (as principal components), for successively maximizing the variance of datasets ([Jolliffe and Cadima, 2016](#)). Briefly, a data matrix of six observable properties for

<https://assignbuster.com/climate-and-human-driven-variability-of-summer-hypoxia-on-a-large-river-dominated-shelf-as-revealed-by-a-hypoxia-index/>

eight cruises ([Supplementary Table 1](#)) was standardized by subtracting the mean and dividing their standard deviation to give a new data matrix, which then was multiplied by the eigenvector to give the variances matrix of six properties for 8 years. We then summed the variances for each year to give the first component. The differences in order of magnitudes of the six properties will not affect the first principal component. The first principal component accounted for 74% of the total variance of the observable properties, and was significantly positively correlated with sectional area ([Supplementary Table 1](#)), mean thickness and the longitude span of the hypoxia area, but significantly negatively correlated with DO minimum, mean bottom DO, and mean DO concentration when $DO < 100\%$, representing the status of the “ severity” of the hypoxia. Therefore, we use the first principal component as a HI to indicate the severity of hypoxia along the transect. The larger the index, the more severe the hypoxia is.

To estimate the contribution of the direct controlling factors to the HI, multiple regression was applied to generate coefficients of the individual variables, which were then standardized by multiplying by the ratio of the standard deviation of each independent variable to the standard deviation of the dependent variable in order to compare them. We used SPSS[®] v. 13 software for PCA and multiple regression, as well as the outlier identification based on a boxplot procedure ([Sim et al., 2005](#)).

Results

DO and Nutrients

The core of the under-saturated water was usually located between 122.5°E and 123.25°E ([Supplementary Figure 2](#)). The area where DO < 42% (~3 mg L⁻¹ in DO concentration) varied in shape over the 8 years, and the largest area was in summer of 2016. The lowest DO saturation of 12% was found in 2017 in the bottom water of station B8 (unpublished data).

Dissolved inorganic nitrogen along the 31°N transect was high in the west and low in the east ([Supplementary Figure 3](#)), due to high concentrations in the Changjiang River plume. The highest concentration was 105 μmol L⁻¹ at station G9 in the summer of 2009. DIN was below 1 μmol L⁻¹ in the outer shelf surface water in the summers of 2006 and 2013.

There were usually two areas of high PO₄³⁻ ([Supplementary Figure 4](#)). One is in Changjiang plume-impacted shelf area, and the other is in the bottom water coming from the outer shelf deep water or Kuroshio subsurface water. The highest concentration of PO₄³⁻ was 2.1 μmol L⁻¹ at station M4-1 during the summer of 2011, while P from the upwelling was up to 1.1 μmol L⁻¹ at the bottom layer of station M4-8 in 2006.

Observable Properties of Hypoxia and the Hypoxia Index

The DO minimum was 69.5 μmol L⁻¹ in 2006 and 42.19 μmol L⁻¹ in 2010, then it jumped to the highest value of 73.44 μmol L⁻¹ in 2013. It then decreased to 26.88 μmol L⁻¹ in 2017 ([Figure 2A](#)).

FIGURE 2

Observable properties of hypoxia including (A) bottom DO, mean DO concentration when $DO < 100\%$ and DO minimum, and (B) sectional area, mean thickness and longitude span used to derive (C) the hypoxia index (HI) during 8 years.

The hypoxia area along the $31^{\circ}N$ transect increased from 0.315 km^2 in 2006 to more than 10 times larger in 2014 and in 2016. The mean thickness also increased from 5.97 m in 2006 to a maximum of 21.9 m in the summer of 2016. The longitude span of the hypoxia area expanded from 2006 to 2016, from less than 95 km to up to 190 km ([Figure 2B](#)).

The HI showed that hypoxia has been becoming more severe ([Figure 2C](#)) in terms of decreasing DO concentration and increasing area, thickness and longitude span from 2006 to 2014, but alleviated a bit after 2014.

Direct Drivers of Hypoxia

Seawater density anomaly of the 8 years showed strong stratification along the $31^{\circ}N$ transect ([Supplementary Figure 5](#)). The density anomaly difference between bottom and surface ($\Delta\sigma$) indicated that strong stratification usually occurred between $122.5^{\circ}E$ and $124^{\circ}E$. The highest was 13.06 kg m^{-3} , at $122.65^{\circ}E$ in summer 2009, while the lowest value of 3.68 kg m^{-3} was from the summer 2006.

τ was the lowest in the 2006, and maximal in 2016 ([Supplementary Table 2](#)). Surface NEP in 2013 was the highest and subsurface NEP was the lowest

<https://assignbuster.com/climate-and-human-driven-variability-of-summer-hypoxia-on-a-large-river-dominated-shelf-as-revealed-by-a-hypoxia-index/>

among the eight cruises, indicating a high export production and respiration rates. Temperature associated with oxygen solubility changed little so we did not present it for the eight cruises.

Discussion

Direct Driver Contribution to Hypoxia Severity

Stratification of the water column plays an important role in hypoxia development ([Zhu et al., 2016](#)). Over the 8 years, the bottom DO was significantly negatively correlated with the density difference ($\Delta\sigma$) between bottom and surface ($r = -0.66$, $p < 0.01$, $n = 47$) ([Figure 3A](#)), indicating the stronger stratification the lower bottom DO. HI was significantly correlated with averaged $\Delta\sigma$ between 122.3 and 124°E ($r = 0.82$, $p < 0.01$, $n = 8$) ([Figure 3B](#)), demonstrating an important role of stratification in regulating hypoxia.

FIGURE 3

Correlations between(A) bottom DO and $\Delta\sigma$ of stations between 122.3–124°E during summers of the 8 years along the transect,(B) HI and mean $\Delta\sigma$ of all stations,(C) HI and residence time (τ) under mixed layer depth (MLD), and(D) hypoxia index (HI) and R in hypoxia area (R_{hy}) for 8 years.

Hypoxia index also significantly correlated with τ ($r = 0.83$, $p < 0.01$, $n = 8$) ([Figure 3C](#)), indicating the longer the water stayed under the MLD, the more severe the hypoxia became, due to insufficient oxygen supply from deep water or the upper mixed layer ([Fennel and Testa, 2018](#)). This is

consistent with previous studies on river-dominated ocean margins ([Rabouille et al., 2008](#)).

In R_{hy} and HI dataset, an outlier (R_{hy} for 2014) was identified. After removing the outlier, the significant correlation between R_{hy} and HI ([Figure 3D](#)) ($r = 0.82$, $p < 0.03$, $n = 7$) showed that community respiration of organic matter by microorganisms such as bacteria ([Chen et al., 2006](#)) was a direct crucial factor that controls the extent of hypoxia. Biogenic particles from the euphotic zone dominated the organic matter as a substrate respired in the aphotic layer in the Changjiang plume-impacted shelf area ([Wang et al., 2016](#)).

In addition, temperature may play a role in determining the respiration rates ([Hopkinson and Smith, 2005](#)), but there was no significant correlation between R_{hy} and bottom temperature, perhaps because temperature changed little in the bottom waters over the eight summers ($21.29 \pm 0.91^\circ\text{C}$). If we use the relationship between pelagic respiration and temperature derived by [Hopkinson and Smith \(2005\)](#), respiration would be $391.9 \pm 12.9 \text{ mmol C m}^{-2} \text{ d}^{-1}$ which varied less than 3% over this narrow temperature range. In fact, respiration may not follow temperature in some estuaries ([Jensen et al., 1990](#) ; [Satta et al., 1996](#)), especially when there is substantial input of external organic matter ([Iriarte et al., 1997](#)). While temperature plays an important role in regulating seasonal respiration rates variation, its influence cannot be distinguished during warm summer periods.

According to a multiple regression analysis, $\Delta\sigma$, τ , and R_{hy} together

explained 95.6% of the total variance of HI ($r = 0.98$, $p = 0.02$, $n = 7$).

<https://assignbuster.com/climate-and-human-driven-variability-of-summer-hypoxia-on-a-large-river-dominated-shelf-as-revealed-by-a-hypoxia-index/>

Area and DO_{\min} are the most common observable properties used to characterize hypoxic zones. The correlation coefficients between the drivers and area or DO_{\min} were lower than for the drivers and HI. The correlations for area and DO_{\min} are 0.79 ($p = 0.23$, $n = 8$) and 0.75 ($p = 0.30$, $n = 8$), respectively, suggesting HI was more robust in describing the severity of hypoxia. The regression equation linking HI to the direct drivers is:

$$HI = (0.06 \pm 0.23) \times \Delta \sigma + (0.09 \pm 0.03) \times \tau + (0.04 \pm 0.02) \times R_{hy} - (5.24 \pm 0.64) \quad (9)$$

The standardized coefficients were 0.06, 0.57, and 0.35, for $\Delta\sigma$, τ , and R_{hy} , respectively. Based on these coefficients, the three factors contributed 5.7, 55.3, and 34.5%, respectively, to the total variance of HI. If we consider $\Delta\sigma$ and τ as two related physical parameters and combine them into one category, then, we could argue that physical stability of the bottom water contributes more than does respiration in forming hypoxia (61.0 versus 34.5%).

Reconstruction of Hypoxia Index for the Past Six Decades

In this section, we reconstruct the HI during the past 60 years, using Eq. 9 and data about direct drivers from field observations and/or correlations between direct and indirect parameters.

To calculate $\Delta\sigma$ over the long-term, we used averaged salinity and temperature of surface and bottom layers in the study area based on historical data from JODC, KODC, and WOA. $\Delta\sigma$ showed an increase at a rate of $0.025 \text{ kg m}^{-3} \text{ yr}^{-1}$ during 1961–1996. After a slight decrease during

1997–2002 it rose after 2003 at a higher rate of $0.047 \text{ kg m}^{-3} \text{ yr}^{-1}$. $\Delta\sigma$ was consistent with the variation of Changjiang River discharge during 1957–2017 ($r = 0.32$, $p < 0.05$, $n = 55$) ([Figure 4A](#)), which freshening the sea surface continuously. Previous studies also suggested a long-term freshening in the northern ECS ([Siswanto et al., 2008](#); [Park et al., 2011](#)). Although the August discharge temporarily decreased by $\sim 10\%$ at June 2003 when the Three Gorges Dam began to impound water ([Chai et al., 2009](#)), it has been increasing before and after 2003 ([Figure 4A](#)). The more stratified conditions allow more severe hypoxia in the study area ([Wei et al., 2007](#); [Zhu et al., 2016](#)).

FIGURE 4

Long-term variation of (A) Changjiang River discharge in August and averaged density difference between bottom and surface ($\Delta\sigma$); (B) averaged temperature (T) and salinity (S) in layer 1 using historical data and water residence time in layer 2 derived from LOICZ model; (C) DIN ($\text{DIN} = \text{NO}_3^- + \text{NO}_2^- + \text{NH}_4^+$) concentration and flux and (D) DIP concentration and flux in the fresh water endmember in Changjiang River; (E) averaged NPP derived from satellite along 31°N transect using the SatCO₂ platform, modeled NPP based on Eq. 11, and modeled respiration in the hypoxia area (R_{hy}); (F) plane area of bottom hypoxia region (<3 and $<2 \text{ mg L}^{-1}$) and DO concentration minimum in historical records ([Gu, 1980](#); [Tian et al., 1993](#); [Li et al., 2002](#); [Quan et al., 2005](#); [Wei et al., 2007](#); [Pei et al., 2009](#); [Li et al., 2018](#)), compared with reconstructed (recon) and PCA-derived hypoxia index

(HI);(G) MEI (multivariate ENSO index) and averaged August SST (COBE-SST2 from NOAA) at 31°N transect in Changjiang plume-impacted shelf area.

The long-term τ has been increasing at a rate of 0.11 d yr^{-1} during 1961–1996, and increased at a much higher rate of 2.52 d yr^{-1} during 2003–2017 ([Figure 4B](#)). This is also associated with the discharge but more closely associated with salinity of the upper mixed layer ([Figure 4B](#)) according to Eq. 6. The fresher the sea surface is, the less water exchange, which allows for longer time for decomposition of organic matter and more consumption of oxygen.

Because respiration data in the hypoxia area are lacking for this long time period, we used NPP based on the significant correlation between NPP and R_{hy} during the 8-year record. As previous studies have found, R_h significantly correlated with NPP ([Hopkinson and Smith, 2005](#)), so we can use NPP to estimate R_{hy} assuming a fixed averaged hypoxic water thickness over longitude to water depth ratio (0.52) derived from the eight cruises, and assuming 50% contribution of sediment to hypoxia:

$$R_{hy} = (1 + 0.5) \times 0.52 \times R_h = 1.5 \times 0.52 \times ((1.60 \pm 0.72) \times \text{NPP} - (59.81 \pm 48.20)) \cdot (10)$$

High NPP was due to higher river DIN flux in recent decades ([Gong et al., 2006](#)). While summer discharge varied relatively little during the past 60 years ($p > 0.05$; [Figure 4A](#)), nutrient concentrations have increased most of the period ([Figures 4C, D](#)) due to anthropogenic activity such as agricultural use of fertilizers and industrial and household pollution ([Wang,](#)

[2006](#); [Li et al., 2007](#), [2016b](#); [Dai et al., 2011](#)). After 2010, both DIN and DIP concentrations and fluxes decreased. Together DIN and DIP fluxes accounted for 56.9% of the variation of NPP, whereas individually DIN and DIP contributed 33.3 and 23.6%, respectively. We can build a relationship between NPP and DIN and DIP fluxes ($p < 0.05$, $n = 12$):

$$\text{NPP} = (4.59 \times 10^{-8} \pm 1.85 \times 10^{-8}) \times \text{F DIN} + (1.27 \times 10^{-6} \pm 7.25 \times 10^{-7}) \times \text{F DIP} + (36.83 \pm 7.96) \quad (11)$$

Then the long-term R_{hy} ([Figure 4E](#)) can be calculated by incorporating Eq. 11 into Eq. 10. Finally the long-term HI variation can be reconstructed with long-term $\Delta\sigma$, τ , and R_{hy} using Eq. 9. The reconstructed HI varied between -4.42 and 2.68 ([Figure 4F](#)), and the standard deviation was from ± 1.95 to ± 3.58 . The highly significant correlation between the reconstructed HI values and PCA-derived HI values (2006–2017) verifies the HI reconstructing model ($r = 0.89$, $p < 0.01$, $n = 8$). The reconstructed HI was highest in 2016 while PCA-derived HI was highest in 2014 ([Figure 4F](#)). This difference could be from the uncertainty of reconstructed HI or the limited sampling locations and periods, suggesting we should be careful when comparing hypoxia severity between specific years, and more historical field data are needed for the verification of HI. However, in summer of 2016 DO was nearly zero north of the 31°N transect during the same cruise, indicating hypoxia was most severe in 2016, consistent with our reconstructed HI. Therefore, we use reconstructed HI for trend analysis.

The hypoxia index showed an overall increase over time accelerating in the last two decades ([Figure 4F](#)). Specifically, HI increased slowly during 1961–
<https://assignbuster.com/climate-and-human-driven-variability-of-summer-hypoxia-on-a-large-river-dominated-shelf-as-revealed-by-a-hypoxia-index/>

1996 (with the rate of 0.042 yr^{-1}), then it increased during 1997–2002 and during 2003–2017 at higher rates (0.222 and 0.298 yr^{-1} , respectively).

Long-term variation of HI is consistent with the general trends of $\Delta\sigma$, τ , and R_{hy} ([Figures 2E, F, 4B](#)), suggesting all three direct drivers also regulate the long-term HI variation.

Climate Change and Human Activities Effects on Inter-Annual Variation of Hypoxia

ENSO and global warming contributed to long-term variation of hypoxia in terms of fluvial discharge freshening Changjiang plume-impacted shelf area, resulting in stronger stratification and longer residence time in the study area ([Figure 5](#)). Indeed, the average salinity in layer 1, were significantly negatively correlated with $\Delta\sigma$ ($r = -0.90$, $p < 0.01$, $n = 53$) and residence time ($r = -0.697$, $p < 0.01$, $n = 53$). The influence of ENSO on the Asian summer monsoon propagates from winter until early summer ([Yang et al., 2018](#)). Warm and cold ENSO phases were represented by MEI, consistent with the surface temperature along the 31°N transect ([Figure 4G](#)). The warm phase (El Niño events) will elevate freshwater discharge of Changjiang River, through affecting precipitation and evaporation in the Changjiang River drainage area during the summer monsoon ([Gong and Ho, 2002](#)). This is supported by the significant positive correlation between MEI (November–December) and August discharge of the next year for Changjiang River during 1955–2017 ([Supplementary Figure 6](#)), similar to the results of [Bai et al. \(2014\)](#). Meltwater from glaciers in the Qinghai-Tibetan Plateau in western China, which has increased due to global warming, is another probable

freshwater source contributing to the runoff increase ([Ding et al., 2006](#); [Siswanto et al., 2008](#)).

FIGURE 5

Schematic diagram of factors controlling estuarine hypoxia. Climate change and anthropogenic activities affect coastal hypoxia by direct drivers through many pathways.

Surface warming would also play a role in regulating stratification in Changjiang plume-impacted shelf area ([Park et al., 2011](#)); the correlation between SST in layer 1 and $\Delta\sigma$ was significant ($r = 0.39$, $p < 0.01$, $n = 53$), suggesting that global warming and ENSO also contributed to the change in stratification in terms of temperature over the past 60 years ([Figure 5](#)). Water masses such as Taiwan Warm Current ([Pei et al., 2009](#)), the Kuroshio Current and Ekman transport could also play a role in regulating the long-term variation in stratification ([Siswanto et al., 2008](#)), but this needs more data and further exploration.

Moreover, the fourth direct driver—oxygen solubility, was closely associated with subsurface water temperature, and rapid warming in the ECS ([Belkin, 2009](#)) will decrease oxygen solubility. Although salinity also affects oxygen solubility, we ignored its effect due to very little variation during the past 60 years ($\sim 0.007 \text{ yr}^{-1}$) and $\sim 1/3$ less significant impact on oxygen solubility in seawater than temperature when they change to the same extent. Our analysis showed bottom water has been warmed up by 4.44°C from 1961 to

2017 [(0.078°C yr⁻¹, much faster than SST increasing at a rate of 0.026°C yr⁻¹ for the whole ECS ([Belkin, 2009](#)) and 0.011°C yr⁻¹ of Kuroshio Current ([Wu et al., 2012](#))]. Since the bottom water mainly comes from Kuroshio Current and will be cooled down and oxygenated each winter, we infer the warming of Kuroshio Current not the bottom water *per se* contributes to the long-term oxygen solubility decrease. Warming rate of 0.011°C yr⁻¹ corresponds to oxygen solubility decrease rate of 0.317 μmol L⁻¹ yr⁻¹ at salinity of 34, then increasing rate of HI will be higher by an additional 0.022 yr⁻¹ (based on the correlation between HI and average bottom DO); namely, warming will make hypoxia ~34% (~10%) more severe by means of decreasing oxygen solubility during 1961–1996 (1997–2017) in the study area.

If we use multiple linear regression again between the long-term reconstructed HI and the three direct drivers, we can differentiate the contribution of $\Delta\sigma$, τ , and R_{hy} to the total variance of HI, as 6.1, 63.6, and 30.2%, respectively. Considering the oxygen solubility as the fourth direct driver, the contribution of four direct drivers would be 4.0, 41.8, 19.8, and 34.4% (5.5, 56.9, 27.0, and 10.6%) during 1961–1996 (1997–2017). Although many studies focus on stratification ([Zhu et al., 2016](#) ; [Zhang et al., 2018](#)) or organic matter-induced respiration ([Chen et al., 2007](#) ; [Zhu et al., 2011](#) ; [Wang et al., 2016](#)) based on short time observations, τ played a more important role than the other three drivers in controlling HI variation in Changjiang plume-impacted shelf area during the past decades. This can be attributed to both advection rate and low DO context of Taiwan Warm

Current ([Qian et al., 2017](#)) or Kuroshio Intermediate Water ([Lui et al., 2014](#)). Other studies also addressed the water residence time as a dominant driver of hypoxia, e. g., in Bellingham Bay ([Wang and Yang, 2015](#)), Gulf of St. Lawrence ([Fennel and Testa, 2018](#)), and comparison study between ECS and Gulf of Mexico ([Rabouille et al., 2008](#)), but few quantified its importance.

Wind plays a less important role in regulating interannual stratification, although eastward Ekman transport, forced by the prevailed southerly or southwesterly wind, may contribute to the eastward extension of the Changjiang River plume ([Siswanto et al., 2008](#)). If we use the wind mixing parameter h/W^3 (h is the water depth, W is the wind speed) with a value of $0.056 \text{ m}^{-2} \text{ s}^3$ ([Xuan et al., 2012](#)) and a highest (mean) wind speed of $5.33 (3.41 \pm 0.86) \text{ m s}^{-1}$ from 1961 to 2017 along the transect during summer, the MLD is estimated to be $8.47 (2.26 \pm 0.035) \text{ m}$. This calculation indicates that wind may influence only the very surface, which is consistent with our 8 years of observations. On the other hand, a typhoon will break the hypoxia, but it would not last long compared with the time scale of weeks to months for hypoxia. Moreover, local hypoxia would persist and even be exacerbated after a typhoon as biological production is enhanced by bring nutrients from bottom to surface water due to the storm mixing ([Ni et al., 2016](#)).

Enhanced respiration was mainly caused by anthropogenic activities releasing nutrients to Changjiang plume-impacted shelf area, resulting in elevated NPP ([Figures 4F, 5](#)). This is consistent with results from the

<https://assignbuster.com/climate-and-human-driven-variability-of-summer-hypoxia-on-a-large-river-dominated-shelf-as-revealed-by-a-hypoxia-index/>

northern Gulf of Mexico ([Rabalais et al., 2014](#)), Chesapeake Bay ([Li et al., 2016a](#)), and other eutrophic coastal waters ([Breitburg et al., 2018](#)).

Increasing temperatures may also lead to higher NPP ([Altieri and Gedan, 2015](#) ; [Li et al., 2016b](#)), but there was no significant correlation between MEI and satellite NPP during 1997–2017, nor was there a significant correlation between SST and NPP. Unlike the eight cruises, the long-term variability in respiration was associated with warming to some extent, as demonstrated by a weak correlation between bottom R_{hy} and temperature ($r = 0.376$, $p < 0.01$, $n = 53$).

Conclusion

Overall, in coastal waters or shelf seas, hypoxia severity in the subsurface water is directly regulated by $\Delta\sigma_t$, R_{hy} , and oxygen solubility. The year-to-year variation in reconstructed HI indicates that summer hypoxia has become more severe since 1960s, because fluvial nutrients concentrations have increased due to increases in anthropogenic activities, and stratification has been stronger and water residence time longer than before. ENSO and global warming may have indirect affects through regulating the fluvial discharge and altering the stratification and residence time by freshening and warming the surface sea, as well as reducing oxygen solubility in the subsurface water. In addition, higher nutrients inputs, as one of the most significant human activities, will enhance primary production and eventually the respiration rates in the subsurface waters; these nutrient fluxes will also be enhanced by elevated discharge. In conclusion, our analyses suggest that climate change accounts for about 80% of the forces

regulating hypoxia in the Changjiang river-impacted shelf while human activity accounts for the remaining 20%.

Data Availability Statement

The original contributions presented in the study are included in the article/[Supplementary Material](#), further inquiries can be directed to the corresponding author/s.

Author Contributions

KW, JC, and W-JC conceived the study. KW, DH, BW, and WF conducted the field studies. W-JC and DK polished the manuscript. KW processed the data. All authors prepared the manuscript.

Funding

This study was funded by National Natural Science Foundation of China (Nos. U1709201, U1609201, 41976154 and 41706120), Long-term Observation and Research Plan in the Changjiang Estuary and East China Sea Project (LORCE). The Scientific Research Fund of the Second Institute of Oceanography, Ministry of Natural Resources, China (No. JT1603). MNR Pilot Project of Early Warning Monitoring of Hypoxia and Ocean Acidification in the Changjiang Estuary (2020–2022).

Conflict of Interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

Acknowledgments

We sincerely thank Maotian Li in East China Normal University for long-term nutrients data at Datong hydrographic station, and Xianqiang He, Feng Zhou, Yan Bai, Qiang Hao, Ping Du, Tao Lian, Xiaobo Ni, Jiliang Xuan, and Dewang Li in Second Institute of Oceanography, Ministry of Nature Resource of China, Willard Moore in University of South Carolina for their helpful comments. We also thank the crews of R/V “ Haijian49,” R/V “ Beidou,” and R/V “ Runjiang” for their help on sampling.

Supplementary Material

The Supplementary Material for this article can be found online at:

[https://www.frontiersin.org/articles/10.3389/fmars.2021.](https://www.frontiersin.org/articles/10.3389/fmars.2021.634184/full#supplementary-material)

[634184/full#supplementary-material](https://www.frontiersin.org/articles/10.3389/fmars.2021.634184/full#supplementary-material)

Footnotes

1. [^ http://www.satco2.com](http://www.satco2.com)
2. [^ http://www.hydroinfo.gov.cn](http://www.hydroinfo.gov.cn)
3. [^ http://www.jodc.go.jp/jodcweb/](http://www.jodc.go.jp/jodcweb/)
4. [^ http://www.nifs.go.kr/kodc/index.kodc](http://www.nifs.go.kr/kodc/index.kodc)
5. [^ https://www.nodc.noaa.gov/OC5/woa13/woa13data.html](https://www.nodc.noaa.gov/OC5/woa13/woa13data.html)
6. [^ http://climexp.knmi.nl](http://climexp.knmi.nl)

References

Altieri, A. H., and Gedan, K. B. (2015). Climate change and dead zones. *Global Change Biol.* 21, 1395–1406. doi: 10.1111/gcb.12754

<https://assignbuster.com/climate-and-human-driven-variability-of-summer-hypoxia-on-a-large-river-dominated-shelf-as-revealed-by-a-hypoxia-index/>

- Bai, Y., He, X., Pan, D., Chen, C. T. A., and Kang, Y. (2014). Summertime Changjiang River plume variation during 1998–2010. *J. Geophys. Res. (C Oceans)* 119, 6238–6257. doi: 10. 1002/2014jc009866
- Belkin, I. M. (2009). Rapid warming of large marine ecosystems. *Prog. Oceanogr.* 81, 207–213. doi: 10. 1016/j. pocean. 2009. 04. 011
- Breitburg, D., Levin, L. A., Oschlies, A., Grégoire, M., Chavez, F. P., and Conley, D. J. (2018). Declining oxygen in the global ocean and coastal waters. *Science* 359: eaam7240.
- Chai, C., Yu, Z., Shen, Z., Song, X., Cao, X., and Yao, Y. (2009). Nutrient characteristics in the Yangtze River Estuary and the adjacent East China Sea before and after impoundment of the Three Gorges Dam. *Sci. Total Environ.* 407, 4687–4695. doi: 10. 1016/j. scitotenv. 2009. 05. 011
- Chen, C. C., Chiang, K. P., Gong, G. C., Shiah, F. K., Tseng, C. M., and Liu, K. K. (2006). Importance of planktonic community respiration on the carbon balance of the East China Sea in summer. *Global Biogeochem. Cycles* 20, 1–14.
- Chen, C. C., Gong, G. C., and Shiah, F. K. (2007). Hypoxia in the East China Sea: one of the largest coastal low -oxygen areas in the world. *Mar. Environ. Res.* 64, 399–408. doi: 10. 1016/j. marenvres. 2007. 01. 007
- Dai, Z., Du, J., Zhang, X., Su, N., and Li, J. (2011). Variation of riverine material loads and environmental consequences on the Changjiang

(Yangtze) Estuary in recent decades (1955-2008). *Environ. Sci. Technol.* 45, 223–227. doi: 10. 1021/es103026a

Ding, Y., Liu, S., Li, J., and Shangguan, D. (2006). The retreat of glaciers in response to recent climate warming in western China. *Annal. Glaciol.* 43, 97–105. doi: 10. 3189/172756406781812005

Fennel, K., and Testa, J. M. (2018). Biogeochemical controls on coastal hypoxia. *Ann. Rev. Mar. Sci.* 11, 105–130. doi: 10. 1146/annurev-marine-010318-095138

Gao, L., Li, D., and Zhang, Y. (2012). Nutrients and particulate organic matter discharged by the Changjiang (Yangtze River): seasonal variations and temporal trends. *J. Geophys. Res. Biogeosci.* 117: G04001.

Gong, D. Y., and Ho, C. H. (2002). Shift in the summer rainfall over the Yangtze River valley in the late 1970s. *Geophys. Res. Lett.* 29: 78-1.

Gong, G. C., Chang, J., Chiang, K. P., Hsiung, T. M., and Hung, C. C. (2006). Reduction of primary production and changing of nutrient ratio in the East China Sea: effect of the three gorges dam? *Geophys. Res. Lett.* 33: L07610.

Gordon, D. C., Boudreau, P. R., Mann, K. H., Ong, J.-E., Silvert, W. L., and Smith, S. V. (1996). *LOICZ Biogeochemical Modelling Guidelines*. Netherlands: LOICZ Core Project.

Grasshoff, K., Kremling, K., and Manfred, E. (1999). eds *Methods of Seawater Analysis*. Weinheim: WILEY-VCH.

<https://assignbuster.com/climate-and-human-driven-variability-of-summer-hypoxia-on-a-large-river-dominated-shelf-as-revealed-by-a-hypoxia-index/>

Gu, H. (1980). The maximum value of dissolved oxygen in its vertical distribution in yellow sea. *Acta Oceanologica Sinica* 2, 70–80.

Guo, X., Xu, B., Burnett, W. C., Wei, Q., Nan, H., and Zhou, S. (2020). Does submarine groundwater discharge contribute to summer hypoxia in the Changjiang (Yangtze) River Estuary? *Sci. Total Environ.* 719: 137450. doi: 10.1016/j.scitotenv.2020.137450

He, X., Pan, D., Bai, Y., Wang, T., Chen, C. T. A., and Zhu, Q. (2017). Recent changes of global ocean transparency observed by SeaWiFS. *Cont. Shelf Res.* 143, 159–166. doi: 10.1016/j.csr.2016.09.011

Holm-Hansen, O., Lorenzen, C. J., Holmes, R. W., and Strickland, J. D. H. (1965). Fluorometric determination of chlorophyll. *ICES J. Mar. Sci.* 30, 3–15. doi: 10.1093/icesjms/30.1.3

Hopkinson, C., and Smith, E. M. (2005). “ Estuarine respiration: an overview of benthic, pelagic, and whole system respiration,” in *Respiration in Aquatic Ecosystems* , eds P. A. D. Giorgio and P. J. L. B. Williams (Oxford: Oxford University Press), 122–146. doi: 10.1093/acprof:oso/9780198527084.003.0008

Huang, R. X., and Russell, S. (1994). Ventilation of the subtropical North Pacific. *J. Phys. Oceanogr.* 24, 2589–2605. doi: 10.1175/1520-0485(1994)024<2589:votsnp>2.0.co;2

Intergovernmental Oceanographic Commission [IOC] (2010). *IAPSO: The International Thermodynamic Equation of Seawater–2010: Calculation and*

<https://assignbuster.com/climate-and-human-driven-variability-of-summer-hypoxia-on-a-large-river-dominated-shelf-as-revealed-by-a-hypoxia-index/>

use of Thermodynamic Properties, Intergovernmental Oceanographic Commission, Manuals and Guides No. 56. New Delhi: IOC.

Iriarte, A., de Madariaga, I., Diez-Garagarza, F., Revilla, M., and Orive, E. (1997). Primary plankton production, respiration and nitrification in a shallow temperate estuary during summer. *J. Exp. Mar. Biol. Ecol.* 208, 127–151. doi: 10.1016/s0022-0981(96)02672-x

Jensen, L. M., Sand-Jensen, K., Marcher, S., and Hansen, M. (1990). Plankton community respiration along a nutrient gradient in a shallow Danish estuary. *Mar. Ecol. Prog. Ser. Oldendorf* 61, 75–85. doi: 10.3354/meps061075

Jolliffe, I. T., and Cadima, J. (2016). Principal component analysis: a review and recent developments. *Phil. Trans. R. Soc. A Mathemat. Phys. Eng. Sci.* 374: 20150202. doi: 10.1098/rsta.2015.0202

Levin, L. A., Ekau, W., Gooday, A. J., Jorissen, F., Middelburg, J. J., and Naqvi, W. (2009). Effects of natural and human-induced hypoxia on coastal benthos. *Biogeosciences* 6, 2063–2098. doi: 10.5194/bg-6-2063-2009

Li, D., Zhang, J., Huang, D., Wu, Y., and Liang, J. (2002). The Oxygen deficit in the Changjiang Estuary. *Sci. China Series D* 32, 686–694.

Li, M., Lee, Y. J., Testa, J. M., Li, Y., Ni, W., and Kemp, W. M. (2016a). What drives interannual variability of hypoxia in Chesapeake Bay: climate forcing versus nutrient loading? *Geophys. Res. Lett.* 43, 2127–2134. doi: 10.1002/2015gl067334

<https://assignbuster.com/climate-and-human-driven-variability-of-summer-hypoxia-on-a-large-river-dominated-shelf-as-revealed-by-a-hypoxia-index/>

Li, M., Wang, H., Li, Y., Ai, W., Hou, L., and Chen, Z. (2016b). Sedimentary BSi and TOC quantifies the degradation of the Changjiang Estuary, China, from river basin alteration and warming SST. *Estuarine Coastal Shelf Sci.* 183, 392–401. doi: 10.1016/j.ecss.2016.08.001

Li, M. T., Xu, K. Q., Watanabe, M., and Chen, Z. Y. (2007). Long-term variations in dissolved silicate, nitrogen, and phosphorus flux from the Yangtze River into the East China Sea and impacts on estuarine ecosystem. *Estuar. Coast. Shelf Sci.* 71, 3–12. doi: 10.1016/j.ecss.2006.08.013

Li, Z., Song, S., Li, C., and Yu, Z. (2018). The sinking of the phytoplankton community and its contribution to seasonal hypoxia in the Changjiang (Yangtze River) estuary and its adjacent waters. *Estuarine Coast. Shelf Sci.* 208, 170–179. doi: 10.1016/j.ecss.2018.05.007

Liu, S. M., Qi, X. H., Li, X., Ye, H. R., Wu, Y., and Ren, J. L. (2016). Nutrient dynamics from the Changjiang (Yangtze River) estuary to the East China Sea. *J. Mar. Syst.* 154, 15–27. doi: 10.1016/j.jmarsys.2015.05.010

Lui, H.-K., Chen, C.-T. A., Lee, J., Bai, Y., and He, X. (2014). Looming hypoxia on outer shelves caused by reduced ventilation in the open oceans: case study of the East China Sea. *Estuar. Coast. Shelf Sci.* 151, 355–360. doi: 10.1016/j.ecss.2014.08.010

Markus Meier, H. E., Väli, G., Naumann, M., Eilola, K., and Frauen, C. (2018). Recently accelerated oxygen consumption rates amplify deoxygenation in the Baltic Sea. *J. Geophys. Res. Oceans* 123, 3227–3240. doi: 10.1029/2017jc013686

<https://assignbuster.com/climate-and-human-driven-variability-of-summer-hypoxia-on-a-large-river-dominated-shelf-as-revealed-by-a-hypoxia-index/>

Matli, V. R. R., Fang, S., Guinness, J., Rabalais, N. N., Craig, J. K., and Obenour, D. R. (2018). Space-Time geostatistical assessment of hypoxia in the Northern Gulf of Mexico. *Environ. Sci. Technol.* 52, 12484–12493. doi: 10.1021/acs.est.8b03474

Meng, J., and Yao, P. (2014). Speciation, bioavailability and preservation of phosphorus in surface sediments of the Changjiang Estuary and adjacent East China Sea inner shelf. *Estuar. Coast. Shelf Sci.* 144, 27–38. doi: 10.1016/j.ecss.2014.04.015

Ni, X., Huang, D., Zeng, D., Zhang, T., Li, H., and Chen, J. (2016). The impact of wind mixing on the variation of bottom dissolved oxygen off the Changjiang Estuary during summer. *J. Mar. Syst.* 154, 122–130. doi: 10.1016/j.jmarsys.2014.11.010

Park, T., Jang, C. J., Jungclaus, J. H., Haak, H., and Park, W. (2011). Effects of the Changjiang river discharge on sea surface warming in the Yellow and East China Seas in summer. *Cont. Shelf Res.* 31, 15–22. doi: 10.1016/j.csr.2010.10.012

Pei, S., Shen, Z., and Laws, E. A. (2009). Nutrient dynamics in the upwelling area of Changjiang (Yangtze River) Estuary. *J. Coast. Res.* 25, 569–580. doi: 10.2112/07-0948.1

Qian, W., Dai, M., Xu, M., Kao, S. J., Chuanjun, D., and Liu, J. (2017). Non-local drivers of the summer hypoxia in the East China Sea off the Changjiang Estuary. *Estuar. Coast. Shelf Sci.* 198, 393–399. doi: 10.1016/j.ecss.2016.08.032

<https://assignbuster.com/climate-and-human-driven-variability-of-summer-hypoxia-on-a-large-river-dominated-shelf-as-revealed-by-a-hypoxia-index/>

Quan, W. M., Shen, X. Q., Han, J. D., and Chen, Y. Q. (2005). Analysis and assessment on eutrophication status and developing trend in Changjiang Estuary and adjacent sea. *Mar. Environ. Sci.* 24, 13–16.

Rabalais, N. N., Cai, W. J., Carstensen, J., Conley, D. J., Fry, B., and Hu, X. (2014). Eutrophication-driven deoxygenation in the coastal ocean. *Oceanography* 27, 172–183. doi: 10. 5670/oceanog. 2014. 21

Rabouille, C., Conley, D. J., Dai, M., Cai, W. J., Chen, C. T. A., and Lansard, B. (2008). Comparison of hypoxia among four river-dominated ocean margins: the Changjiang(Yangtze), Mississippi, Pearl, and Rhone rivers. *Cont. Shelf Res.* 28, 1527–1537. doi: 10. 1016/j. csr. 2008. 01. 020

Resplandy, L. (2018). *Will Ocean zones With Low Oxygen Levels Expand or Shrink?* Berlin: Nature Publishing Group.

Satta, M. P., Agustí, S., Mura, M. P., Vaqué, D., and Duarte, C. M. (1996). Microplankton respiration and net community metabolism on the NW Mediterranean coast. *Aquat. Microb. Ecol.* 10, 165–172. doi: 10. 3354/ame010165

Scavia, D., Justic, D., Obenour, D. R., Craig, K., and Wang, L. (2018). Hypoxic volume is more responsive than hypoxic area to nutrient load reductions in the northern Gulf of Mexico—and it matters to fish and fisheries. *Environ. Res. Lett.* 14: 024012. doi: 10. 1088/1748-9326/aaf938

Sim, C. H., Gan, F. F., and Chang, T. C. (2005). Outlier labeling with boxplot procedures. *J. Am. Statist. Assoc.* 100, 642–652. doi: 10.

1198/016214504000001466

Siswanto, E., Nakata, H., Matsuoka, Y., Tanaka, K., Kiyomoto, Y., and Okamura, K. (2008). The long-term freshening and nutrient increases in summer surface water in the northern East China Sea in relation to Changjiang discharge variation. *J. Geophys. Res. Oceans* 113: C10.

Song, G., Liu, S., Zhu, Z., Zhai, W., Zhu, C., and Zhang, J. (2016). Sediment oxygen consumption and benthic organic carbon mineralization on the continental shelves of the East China Sea and the Yellow Sea. *Deep Sea Res. Part II: Top. Stud. Oceanography* 124, 53–63. doi: 10.1016/j.dsr2.2015.04.012

Taylor, J. R. (1997). (Ed). *An Introduction to Error Analysis*. Mill Valley, CA: University Science Books.

Tian, R. C., Hu, F. X., and Martin, J. M. (1993). Summer nutrient fronts in the Changjiang (Yangtze River) Estuary. *Estuar. Coast. Shelf Sci.* 37, 27–41. doi: 10.1006/ecss.1993.1039

Vaquer-Sunyer, R., and Duarte, C. M. (2008). Thresholds of hypoxia for marine biodiversity. *Proc. Natl. Acad. Sci. U S A.* 105, 15452–15457. doi: 10.1073/pnas.0803833105

<https://assignbuster.com/climate-and-human-driven-variability-of-summer-hypoxia-on-a-large-river-dominated-shelf-as-revealed-by-a-hypoxia-index/>

Wang, B. D. (2006). Cultural eutrophication in the Changjiang (Yangtze River) plume: history and perspective. *Estuar. Coast. Shelf Sci.* 69, 471–477. doi: 10. 1016/j. ecss. 2006. 05. 010

Wang, H., Dai, M., Liu, J., Kao, S. J., Zhang, C., Cai, W. J., et al. (2016). Eutrophication-driven hypoxia in the East China Sea off the Changjiang Estuary. *Environ. Sci. Technol.* 50, 2255–2263. doi: 10. 1021/acs. est. 5b06211

Wang, T., and Yang, Z. (2015). Understanding the flushing capability of Bellingham Bay and its implication on bottom water hypoxia. *Estuar. Coast. Shelf Sci.* 165, 279–290. doi: 10. 1016/j. ecss. 2015. 04. 010

Wang, X., Baskaran, M., Su, K., and Du, J. (2018). The important role of submarine groundwater discharge (SGD) to derive nutrient fluxes into River dominated Ocean Margins–The East China Sea. *Mar. Chem.* 204, 121–132. doi: 10. 1016/j. marchem. 2018. 05. 010

Wei, H., He, Y., Li, Q., Liu, Z., and Wang, H. (2007). Summer hypoxia adjacent to the Changjiang Estuary. *J. Mar. Syst.* 67, 292–303. doi: 10. 1016/j. jmarsys. 2006. 04. 014

Wei, Q., Wang, B., Yu, Z., Chen, J., and Xue, L. (2017). Mechanisms leading to the frequent occurrences of hypoxia and a preliminary analysis of the associated acidification off the Changjiang estuary in summer. *Sci. China Earth Sci.* 60, 360–381. doi: 10. 1007/s11430-015-5542-8

<https://assignbuster.com/climate-and-human-driven-variability-of-summer-hypoxia-on-a-large-river-dominated-shelf-as-revealed-by-a-hypoxia-index/>

Wolter, K., and Timlin, M. S. (1998). Measuring the strength of ENSO events: how does 1997/98 rank? *Weather* 53, 315–324. doi: 10. 1002/j. 1477-8696. 1998. tb06408. x

Wu, L., Cai, W., Zhang, L., Nakamura, H., Timmermann, A., and Joyce, T. (2012). Enhanced warming over the global subtropical western boundary currents. *Nat. Climate Change* 2, 161–166. doi: 10. 1038/nclimate1353

Xuan, J. L., Huang, D., Zhou, F., Zhu, X. H., and Fan, X. (2012). The role of wind on the detachment of low salinity water in the Changjiang Estuary in summer. *J. Geophys. Res. Oceans* 117: C10004.

Yang, Z., Chen, J., Chen, M., Ran, L., Li, H., and Lin, P. (2018). Sources and transformations of nitrogen in the South China Sea: insights from nitrogen isotopes. *J. Oceanogr.* 74, 101–113. doi: 10. 1007/s10872-017-0443-z

Zhang, W., Wu, H., Hetland, R. D., and Zhu, Z. (2019). On mechanisms controlling the seasonal hypoxia hot spots off the Changjiang River Estuary. *J. Geophys. Res. Oceans* 124, 8683–8700. doi: 10. 1029/2019jc015322

Zhang, W., Wu, H., and Zhu, Z. (2018). Transient hypoxia extent off changjiang river estuary due to mobile Changjiang River Plume. *J. Geophys. Res. Oceans* 123, 9196–9211. doi: 10. 1029/2018jc014596

Zhou, F., Chai, F., Huang, D., Xue, H., Chen, J., and Xiu, P. (2017). Investigation of hypoxia off the Changjiang Estuary using a coupled model of ROMS-CoSiNE. *Prog. Oceanogr.* 159, 237–254. doi: 10. 1016/j. pocean. 2017. 10. 008

<https://assignbuster.com/climate-and-human-driven-variability-of-summer-hypoxia-on-a-large-river-dominated-shelf-as-revealed-by-a-hypoxia-index/>

Zhu, J., Zhu, Z., Lin, J., Wu, H., and Zhang, J. (2016). Distribution of hypoxia and pycnocline off the Changjiang Estuary. *China. J. Mar. Syst.* 154, 28–40. doi: 10. 1016/j. jmarsys. 2015. 05. 002

Zhu, Z. Y., Zhang, J., Wu, Y., Zhang, Y. Y., Lin, J., and Liu, S. M. (2011). Hypoxia off the Changjiang (Yangtze River) Estuary: oxygen depletion and organic matter decomposition. *Mar. Chem.* 125, 108–116. doi: 10. 1016/j. marchem. 2011. 03. 005